| 1 | Temporal variability of reactive iron over the Gulf of Alaska Shelf |
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1 Abstract

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3 The Gulf of Alaska (GoA) shelf is a highly productive regime bordering the nitrate-rich, iron (Fe) limited waters of the central GoA. The exchange between nitrate-limited, Fe-replete coastal 4 waters and nitrate-rich, Fe-deplete offshore waters, amplified by mesoscale eddies, is key to the 5 productivity of the region. Previous summer field studies have observed the partitioning of Fe in 6 7 the coastal GoA as being heavily dominated by the particulate phase due to the high suspended particulate loads carried by glacial rivers into these coastal waters. Here we present new physico-8 chemical iron data and nutrient data from the continental shelf of the GoA during spring and late 9 summer, 2011. The late summer data along the Seward Line showed variable surface dissolved 10 iron (DFe) concentrations (0.052 nM offshore to 4.87 nM inshore), within the range of previous 11 12 observations. Relative to available surface nitrate, DFe was in excess (at Fe:C =50 µmol:mol) inshore, and deficient (at $Fe:C = 20 \mu mol:mol$) offshore. Summer surface total dissolvable iron 13 (TDFe, acidified unfiltered samples) was dominated by the acid-labile particulate fraction over 14 the shelf (with a median contribution of only 3% by DFe), supporting previously observed Fe 15 partitioning in the GoA. In contrast, our spring data from southeast GoA showed TDFe 16 17 differently partitioned, with surface DFe (0.28 - 4.91 nM) accounting on average for a much 18 higher fraction (~ 25%) of the TDFe pool. Spring surface DFe was insufficient relative to available nitrate over much of the surveyed region (at $Fe:C = 50 \mu mol:mol$). Organic Fe-binding 19 20 ligand data reveal excess concentrations of ligands in both spring and summer, indicating incomplete titration by Fe. Excess concentrations of an especially strong-binding ligand class in 21 22 spring surface waters may reflect *in-situ* ligand production. Due to anomalous spring conditions in 2011, river flow and phytoplankton biomass during our spring sampling were lower than the 23 expected May average. We argue our samples are likely more representative of early spring pre-24 bloom conditions, providing an idea of the possible physico-chemical partitioning of Fe in 25 26 coastal GoA waters relevant to initial spring bloom dynamics.

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1 1. Introduction

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3 The essential micro-nutrient iron (Fe) regulates primary productivity over large areas of the open ocean including the Gulf of Alaska (GoA) high-nutrient, lower than expected chlorophyll 4 (HNLC) region (e.g. Martin et al., 1989; Boyd et al., 2004). Additionally, Fe availability 5 influences algal community composition, which transitions from a diatom-dominated community 6 7 at high concentrations of Fe to a system dominated by small phytoplankton at low Fe concentrations (Morel et al., 1991; Landry et al., 1997). The biological regulatory effect of Fe 8 9 can also influence coastal waters (Hutchins et al., 1998; Johnson et al., 2001). As a consequence, it is of interest to understand the factors that affect the distribution of biologically accessible Fe 10 in surface waters. Physical, chemical, and biological processes transform Fe in seawater by 11 altering its distribution within various pools, and thus its biological availability. Truly soluble (< 12 $0.02 \,\mu\text{m}$) inorganic Fe, although it appears to be easily accessible to phytoplankton (Maldonado 13 and Price, 2000), is found at exceedingly low concentrations due to its extreme low solubility in 14 oxygenated seawater. Organic complexation increases the solubility of Fe in seawater (Liu and 15 16 Millero, 2002), and >99% of dissolved (< 0.4 μ m) Fe (DFe) is organically complexed by strong Fe-binding ligands that can allow for elevated DFe concentrations ([DFe]) (Buck et al., 2007; 17 18 Bundy et al., 2014). However, the biological availability of these complexes has been shown to be variable (e.g. Maldonado and Price, 1999; Rijkenberg et al., 2006). In addition, a fraction of 19 20 the suspended particulate Fe pool may be mobilized into dissolved or soluble phases over short time scales, rendering this labile fraction potentially available to phytoplankton (Johnson et al., 21 22 1999). Together DFe and suspended labile particulate Fe represent reactive species that are potentially significant to biological uptake. 23

24 Surface waters exhibit pronounced inshore-offshore gradients in DFe and suspended 25 particulate Fe, with inshore Fe concentrations enhanced by up to 2-3 orders of magnitude (or greater), as has been observed in the GoA (Wu et al., 2009; Lippiat et al., 2010a). The gradients 26 result from the proximity of inshore waters to terrestrial Fe sources (aeolian, fluvial and 27 sedimentary). Atmospheric deposition, via episodic dust events from Asia and North America, 28 29 has been traditionally thought as the main mechanism for transporting Fe to the central GoA (Boyd et al., 1998; Moore et al., 2002). Recent studies (Johnson et al., 2005, Lam et al., 2006, 30 31 Cullen et al., 2009; Lippiatt et al., 2011; Brown et al., 2012) suggest fluxes of Fe from coastal

waters to the central GoA could be as important as atmospheric deposition. For example, 1 2 anticyclonic mesoscale eddies in the GoA that propagate from the inner shelf westward into the 3 basin (Ladd et al., 2005), are able to transport Fe-rich waters offshore (Johnson et al., 2005; Lippiat et al., 2011; Brown et al., 2012), and it has been calculated that the flux of reactive Fe 4 from mesoscale eddy activity in the GoA is in the same order of magnitude as atmospheric flux 5 6 (Brown et al., 2012). Lam et al. (2006) presented evidence of lateral advection of suspended 7 particulate Fe from the continental shelf into the remote HNLC region at depths of 0-300 m, and Lam and Bishop (2008) calculated the flux of biologically available Fe to be within the same 8 9 order of magnitude for these two sources by assuming higher lability for sedimentary Fe as 10 compared to dust. Additional cross-shelf Fe transport via the California Undercurrent, downwelling, and/or tidal currents could enhance the importance of coastal input relative to 11 12 atmospheric deposition, as suggested by Cullen et al. (2009).

Input of Fe to the GoA is highly seasonal. In the northern GoA prevailing winds stimulate 13 downwelling from fall through spring (Stabeno et al., 2004; Weingartner et al., 2005), and likely 14 contribute to episodic offshore flow of Fe in bottom layers. During May-October, weaker and 15 16 variable winds induce intermittent upwelling along the coast (Stabeno et al., 2004). Fresh water runoff is at its peak between June and September, and higher concentrations of Fe have been 17 18 observed in July as compared to May in the inner GoA shelf (Wu et al., 2009). The various glaciers along the mountainous GoA coastline cause rapid erosion and contribute to the high 19 20 suspended sediment (Christensen et al., 2000) and Fe loads (Lippiatt et al., 2010a) carried by rivers. The rapid physical weathering of glaciers produces particles with a low labile Fe 21 22 component (Lippiatt et al., 2010a) as compared to more labile Fe-coatings on resuspended particles that have been exposed to chemical reduction and re-oxidation processes at the 23 24 sediment-water interface (e.g. Hurst et al., 2010). Dust storms generated by episodic gap wind events (Ladd and Chen, 2015; Ladd et al., 2015) can deposit glacial flower with high Fe loadings 25 (Crusius et al., 2011) over the shelf and offshore waters from exposed river beds of glacial rivers. 26 These events occur most frequently in mid-autumn (Ladd and Chen, 2015; Ladd et al., 2015) 27 when low river levels and snow cover expose the riverbed sediment (Crusius et al., 2011 and 28 29 references therein). The concentration of dissolved organic Fe-binding ligands likely caps the amount of Fe from various inputs that contribute to the DFe pool, as observed in other river 30 31 influenced coastal systems (Buck et al., 2007), but inorganic colloidal Fe is potentially an

important component of the dissolved Fe pool in the high particle regime of the inner GoA shelf
during summer. Because most fluvial input of Fe is confined within the buoyancy driven Alaska
Coastal Current (ACC) (Figure 1), the ACC has the potential to act as Fe reservoir, and a vehicle
for alongshore transport (Wu et al., 2009).

We present Fe and ligand data from the continental shelf of southeast GoA during spring, and 5 the western GoA during late summer, 2011 (Figure 1), collected as part of the Gulf of Alaska 6 Integrated Ecosystem Research Program (GOA-IERP). Dissolved and particulate samples were 7 collected from surface waters and from depth profiles. Samples were processed to determine the 8 partitioning of Fe into various size classes and chemical species in order to show how the 9 physico-chemical speciation of Fe changes as a function of season and location in the Gulf of 10 Alaska, reflecting differences in Fe inputs. The various fractions of Fe discussed in this paper are 11 defined in Table 1. The study area borders the central HNLC GoA region and, considering the 12 recently described importance of cross-shelf processes in delivering Fe to the central GoA, the 13 seasonal partitioning of Fe over the shelf is likely important in determining the offshore flux of 14 biologically available Fe to the subarctic north Pacific. 15

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- 17 **2. Materials and Methods**
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19 2.1. Sampling Region

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Observations were made in spring 2011 (30 April – 21 May) onboard the R/V Thomas G. 21 22 Thompson in the eastern GoA, and in late summer 2011 (14 - 20 September) onboard the M/V*Tiglax* in the western GoA. During spring, seven transects were opportunistically sampled from 23 24 Chatham Strait to Kayak Island (Figure 1), and vertical profiles were collected from a subset of 25 GOA-IERP stations along three cross-shelf lines (South East Line A (SEA), South East Line G (SEG), and Yakutat Bay Line C (YBC)) (Figure 1). Stations along the SEA line were sampled on 26 5 May, those along the SEG line were sampled twice (on 7 May and 17 May), and the YBC 27 stations were sampled on 11 May. In late summer, surface samples were collected along the 28 29 Seward Line on 16, 17, and 19 September, and in Prince William Sound on 15 September 30 (Figure 1).

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1 2.2 Sampling Protocols

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3 Sampling was conducted using trace-metal clean procedures recognized by the International GEOTRACES Program (http://www.geotraces.org/science/intercalibration/222-4 sampling-and-samplehandling-protocols-for-geotraces-cruises). Underway surface (~1 m) 5 sampling was accomplished using a towed trace metal clean pump system (Bruland et al., 2005; 6 Aguilar-Islas and Bruland, 2006). This system is composed of a PTFE diaphragm pump 7 (Wilden) and PFA Teflon[™] tubing mounted onto a custom-made PVC vane which is attached to 8 a 20 kg PVC torpedo that hangs 1 m below the vane. The vane-torpedo assembly is deployed 9 from the side of the ship and towed away from the hull at speeds of 7-10 knots, depending on sea 10 conditions. Surface samples from the towed system were filtered in-line through a 0.2 µm pore 11 12 Supor Acropak 200 filter capsule (Pall Corporation) which was rinsed with at least 10 L of seawater prior to use. Filtered and unfiltered samples for Fe analysis were collected into pre-13 cleaned low density polyethylene (LDPE) bottles, acidified to pH ~1.8 (Optima grade HCl, 14 Fisher Scientific) (2 ml concentrated HCl/L) upon return to the laboratory, and analyzed after 15 16 one month of storage at room temperature. Filtered samples for Fe-binding organic ligands were collected in fluorinated polyethylene acid-cleaned bottles and stored at -20 °C until analysis 17 (Buck et al., 2012). Filtered samples for macro-nutrient determination were collected in 60 mL 18 acid washed, high-density polyethylene bottles after three rinses. Samples were frozen (-20 °C) 19 20 and shipped to the laboratory for analysis.

Depth profiles for Fe were collected using University of Alaska Fairbanks (UAF) vanes (Wu 21 22 2007, Wu et al., 2009). UAF vanes are a modified version of the Moored In-Situ Trace Element Serial Sampler (MITESS, Bell et al., 2002) consisting of a single MITESS module attached to a 23 24 2-L polyethylene square bottle. The assembly was mounted onto a plastic vane that was attached to the ship's hydrowire. The bottom-most vane was positioned 10 m above the ship's rosette 25 26 system. The MITESS modules were programmed to open/close bottles simultaneously after bottles sat for at least 10 minutes at the target depths. During sample collection, the vane portion 27 of the assembly positions the sampling bottle upstream of the hydrowire, allowing for the 28 29 collection of uncontaminated seawater samples. Upon retrieval, the 2L bottles were stored in the refrigerator until processed (0-6hrs). Seawater was vacuum-filtered through acid-cleaned 47 mm 30 31 diameter 0.4 µm track-etched polycarbonate membranes (Nuclepore, Whatman), and the filtrate

was collected in the same bottle types as for surface samples. The membranes were folded into
eighths, placed into pre-cleaned 7 mL polypropylene vials and stored at -20 °C. Blank filter
membranes were treated in the same manner as samples, but using Milli-Q water (18.2 MΩ cm)
instead of seawater.

Sample processing during spring (*R/V Thompson*) was carried out within a plastic 5 6 enclosure under HEPA filtered air (ISO Class 5) in the ship's main laboratory. During late 7 summer (M/V Tiglax), due to on-board laboratory space limitations, depth profiles were not 8 collected and surface in-line unfiltered and filtered samples were collected under plastic bells on 9 the deck of the ship, taking care to avoid contamination. Plastic bells were made with lidded plastic containers outfitted with a bulkhead tubing connector to interface with the tubing coming 10 from the surface pump system into the inside of the container. The plastic container was large 11 enough to accommodate maneuverability during sampling. The lidded end of the bell was kept 12 13 closed in between samplings. Two bells were constructed, one for filtered and the other for unfiltered sample collection. Table 2 summarizes sampling operations during 2011. 14

Conductivity, temperature and depth (CTD) data for profiles on the *R/V Thomas G*. 15 Thompson, were obtained from a SeaBird SBE 911 plus CTD sensor within a 24-bottle rosette 16 17 system, chlorophyll fluorescence from a WetLabs ECO chlorophyll fluorometer with factory calibration, and light transmission was measured with a Wet Labs CSTAR Transmissometer. 18 19 Underway salinity, temperature, and fluorescence data was obtained from the underway system of the *R/V Thomas G. Thompson* during spring. The fluorometer for the underway system was 20 21 not calibrated at sea; therefore, the raw data represents a relative measure of chlorophyll fluorescence levels at the surface. During late summer on the *M/V Tiglax*, underway salinity and 22 23 temperature were obtained from a Midi Hydro-Bios MultiNet equipped with a Hydro-Bios CT-Set sensor. The towed trace metal clean pump system was deployed (recovered) immediately 24 after (before) MultiNet recovery (deployment), and the obtained temperature and salinity were 25 from within 500 m from the sample collected for Fe analyses. 26

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28 2.3. Suspended Particle Processing

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Filters containing suspended particulate samples were thawed, leached and digested upon
return to the laboratory. Optima grade acids (Fisher Scientific) were employed during leach and

digestion protocols. A modified leach technique based on the one described by Berger et al., 1 2 (2008) was employed. Thaved folded filters were kept in their 7 mL vials, and 1 mL of 25 % 3 acetic acid with 0.02 M hydroxylamine hydrochloride (NH₂OH • HCl) was added to each vial. The vials were then placed in a water bath, heated to 90-95 °C for 10 min, and left to cool 4 5 gradually down to 30 °C for the duration of the 2 h leach. The heating step releases intracellular Fe while the reducing agent accesses the portion of Fe associated with oxyhydroxide coatings 6 7 (Berger et al., 2008). Leachates were transferred into 10 mL PTFE beakers (Jensen Inert 8 Products) and heated to near dryness. After the beakers cooled, $100 \,\mu$ L of 1 M HNO₃ was added 9 and heated. The solution was transferred into acid-clean 7 mL polyethylene vials by rinsing the beaker 3 times with 1 M HNO₃ (1 mL per rinse) and the solution was stored until analysis at 10 11 room temperature. Each leached folded filter was transferred to a 15 mL Teflon vial (Savillex 12 Corp.) and a 4-step digestion, modified from Morton et al., (2013), was carried out: 1) 1 mL of concentrated HNO₃ was added and allowed to reflux for 12 h at 130 °C, then the solution was 13 taken down to near dryness; 2) a mixture of concentrated HNO₃ (1 mL), concentrated HCl (500 14 µL) and concentrated HF (100 µL) was added and allowed to reflux for 12 h at 130 °C, then 15 16 taken down to near dryness; 3) concentrated HNO₃ (1 mL) and HCl (500 μ L) were added and 17 left to reflux for 12 h at 130 °C, then taken down to near dryness; 4) 1 mL of concentrated HNO₃ was added and refluxed for 30 min at 130 °C. Once cooled, the digested solution was transferred 18 to a LDPE 30 mL bottle, and the vial was rinsed 4 times with 0.1% HNO₃ to vield a final 5% 19 20 HNO₃ solution. Blank filters were leached and digested in the same manner as samples. Blanks for the leaching (0. 70 ppb \pm 0.1 ppb (n = 7)) and digestion (0.25ppb \pm 0. 07 ppb (n = 8)) 21 22 procedures represented 0 - 13% and 1 - 29% of the samples, respectively. 23 24 2.4. Analytical methods 25 2.4.1. Iron determination 26 27 Analyses were carried out at the University of Alaska Fairbanks on an Element 2 (Thermo-28 Finnigan) inductively couple plasma mass spectrometer. Seawater DFe and TDFe were 29 quantified by isotope dilution with high resolution inductively coupled plasma mass 30

31 spectrometry (HR-ICP-MS) detection after Mg(OH)₂ co-precipitation (Wu and Boyle 1998).

Briefly, 1.6 mL of filtered seawater sample and the isotope spike are mixed in a 2 mL vial 1 2 followed by the addition of ammonium hydroxide. The vial is then centrifuged, the supernatant is 3 decanted, and the precipitate is diluted in 4% HNO₃ (Optima grade, Fisher Scientific). 4 Procedural blanks were determined using 50 μ l of seawater with a known low (<0.1 nM) [DFe] and were on average 0.05 nM (n = 8), and the detection limit (3σ) was 6 pM. Accuracy was 5 determined from the analysis of the reference sample SAFE D2 (0.923 ± 0.008 nM; n = 13) 6 7 which was within the latest community consensus value (SAFe D2 = 0.933 ± 0.023 nmol/kg) reported in May, 2013 (www.geotraces.org). Suspended LPFe and RPFe were determined by 8 9 direct injection into the mass spectrometer as described by Aguilar-Islas et al., (2013) using a calibration curve and Ga as an internal standard. Values are reported as nmol of Fe per L of 10 11 filtered seawater (nM).

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13 2.4.2. Organic Fe-binding ligand determination

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Dissolved Fe-binding ligand concentrations and conditional stability constants were determined 15 by competitive ligand exchange-adsorptive cathodic stripping voltammetry (CLE-ACSV) with 16 17 the added competitive ligand salicylaldoxime (Rue and Bruland, 1995; Buck et al., 2007; 2012). 18 Frozen seawater samples were thawed at room temperature, shaken, and distributed into 12 Teflon lidded vials (Savillex) in 10 mL aliquots. A borate-ammonium buffer was used to 19 maintain pH of 8.2 (NBS; Ellwood and van den Berg, 2001) in all vials. Dissolved Fe additions 20 employed in the titrations were +0, 0, 0.25, 0.5, 0.75, 1, 1.5, 2, 2.5, 3.5, 5, 7.5 nM. Following 21 22 two or more hours of equilibration between the dissolved Fe additions and natural ligands in the sample, $25 \,\mu$ M of the competitive ligand salicylaldoxime (SA) was added to each vial and the 23 resulting Fe-SA complex was then determined by ACSV on a BAS*i* controlled growth mercury 24 25 electrode (Buck et al., 2007; 2012). Resulting titration data were interpreted using a combination of van den Berg/Ruzic and Scatchard linearization approaches, and reported results represent the 26 27 average outputs from the two linearizations (Buck et al., 2007; 2012). Up to two ligand classes were determined from each individual titration, while the notation for a given sample's ligand 28 classes was defined by the conditional stability constant determined for each class: Ligands with 29 $\log K_{FeL_i,Fe\ell}^{cond} > 12$ are defined as L₁, L₂ for $\log K_{FeL_i,Fe\ell}^{cond} = 11 - 12$, and L₃ for $\log K_{FeL_i,Fe\ell}^{cond} = 10 - 11$ 30 (Gledhill and Buck, 2012). 31

2 2.4.3. Macro-nutrient determination

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Concentrations of nitrate, nitrite, ammonium, phosphate and silicic acid were determined using a
combination of analytical components from Alpkem, Perstorp and Technicon. The WOCEJGOFS standardization and analysis procedures specified by Gordon et al. (1994) were closely
followed including reagent preparation, calibration of labware, preparation of primary and
secondary standards, and corrections for blanks and refractive index. Concentrations of
secondary standards were verified using commercial standards from OSIL.

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11 2.4.4 Iron Fractions

12 Four operationally defined Fe species were measured: 1) Dissolved Fe (DFe) is the filtered fraction of iron that passes through the inline 0.2 µm filter capsule (tow fish) or the 0.4 µm filter 13 membrane (UAF vanes); 2) Total dissolvable Fe (TDFe) is the fraction of iron obtained from 14 unfiltered samples acidified to pH ~1.8. This was only obtained from tow fish samples; the 15 particulate Fe fraction (retained by the 0.4 µm filter membrane) was obtained from vertical 16 17 profile samples, and 3) Leachable particulate Fe (LPFe) is the fraction of particulate Fe 18 solubilized by an acetic acid leach; while 4) Refractory particulate Fe (RPFe) is the strong acid digested particulate Fe fraction. Based on these measured species, three calculated fractions were 19 20 obtained: 1) Particulate Fe is the sum of LPFe and RPFe; 2) Acid-labile particulate Fe (ALPFe) is the fraction of particulate Fe that dissolves when an unfiltered sample is acidified to pH 1.8 21 (TDFe – DFe); and 3) Reactive Fe is the fraction that is kinetically reactive and potentially 22 23 available to biology in short timescales of days to weeks, defined by Lippiat et al., (2010a) as DFe + LPFe, and here also defined by TDFe. When referring to "labile particulate Fe" we 24 25 include both ALPFe and LPFe fractions. These definitions of various Fe fractions are summarized in Table 1. 26

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28 **3. Results**

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The upwelling index (NOAA Bakun Index) for the region indicated downwelling-favorable
winds over most of both sampling periods, with short episodes of relaxation and weakly

upwelling winds during both seasons. As expected, river discharge during our spring sampling 1 2 was low, with an average daily discharge for the Alsek and Copper rivers of 376 m³/sec and 837 3 m^{3} /sec, respectively, as compared to the average daily 2011 peak discharge flow in July and August of 2237 m³/sec for the Alsek River and 4546 m³/sec for the Copper River (USGS 4 surface-water daily statistics data, 2011). However, river discharge during our spring sampling 5 was also ~ 1/3 lower relative to the same sampling period in previous years (2004-2010; 573) 6 7 m³/sec for the Alsek River and 1232 m³/sec for the Copper River) (available USGS surfacewater daily statistics data). This lower discharge was in part related to the delay in the onset of 8 the spring freshet in 2011 (USGS surface-water daily statistics data) as compared to previous 9 10 years (2004-2010; data not shown).

11 Relative to climatological May values (Waite and Mueter, 2013), the biomass present during 12 our May 2011 sampling was anomalously low (Strom and Fredrickson, 2015), and this 13 anomalously low production persisted throughout the spring of 2011 (Stabeno et al., 2015). The 14 phytoplankton community in May 2011 is described by Strom and Fredrickson (2015) as 15 consisting mainly of picophytoplankton (< 2 μ m) cells that exhibited low specific growth rates 16 and acclimation to low light conditions.

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18 3.1 Spring Surface Transects

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20 Temperature (T), salinity (S), nitrate plus nitrite (N+N), fluorescence, DFe and TDFe along transects are presented in Figures 2a-f. Surface T ranged from ~ 6 °C to 9 °C with the lowest T 21 22 values near shore in the vicinity of Chatham Strait and Cross Sound. Higher salinity Alaska Gyre 23 water (S = 32.6-32.8, Musgrave et al., 1992) was not encountered during spring. Sampled surface 24 waters ranged in S from 30.95 to 32.25 with the lowest near Yakutat Bay and Kayak Island, and 25 the highest offshore of Cross Sound and Yakutat Bay. In general the concentration of N+N had 26 not been depleted, and concentration ranged from 0.18 to12.27 µM (Figure 2d). Samples with 27 low N+N (< 1 µM) were collected in the vicinity of Cross Sound and Kayak Island. Chlorophyll fluorescence values from the underway sensor varied from ~ 0.02 to 0.2V, with a narrow area of 28 higher values in late May near Chatham Strait. In situ Chl a and satellite data (Strom and 29 Fredrickson, 2015; Stabeno et al., 2015; Waite and Mueter, 2013) indicated an anomalously low 30 spring bloom during 2011, especially in the eastern GoA. The concentration of DFe was on 31

average 1.52 ± 1.04 nM, and ranged from 0.28 to 4.91 nM. The highest concentrations were 1 2 observed near Cross Sound and Kayak Island, and the lowest concentrations were observed 3 offshore. There was no clear trend for DFe over the narrow salinity range sampled, although the 4 freshest water collected (S = 30.95) had an elevated [DFe] of 4.50 nM, and the lowest surface DFe (0.28 nM) was measured in offshore higher salinity water (S = 32.11). The concentration of 5 TDFe ranged over nearly 3 orders of magnitude, from 0.72 to 260 nM, with a median value of 6 7 4.73 nM and a few samples near Cross Sound, Yakutat Bay, and Kayak Island containing TDFe in excess of 100 nM. The partitioning of TDFe between ALPFe and DFe was highly variable, 8 9 with DFe accounting for ~ 1.5% to 80% of the TDFe pool (~25% on average) over the sampled 10 area. A subset of spring surface samples analyzed for dissolved Fe-binding organic ligands showed 11 12 two ligand classes (stronger (L_1) and relatively weaker (L_2)) present in all samples (Table 3). The

13 L₁-type ligands had high conditional stability constants (log $K_{FeL_1,Fet}^{cond} = 12.2 \pm 0.17$ to 13.0 ± 0.05 ;

14 Table 3) and L_1 concentrations were in excess of DFe, dominating the ligand pool. The

15 conditional stability constants for L_1 in the spring surface samples were the highest of any

samples analyzed in this Gulf of Alaska dataset (Table 3).

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18 3.2 Spring Depth Profiles

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Vertical profiles of T, S, DFe, beam transmission (Beam c), Chl a, LPFe and RPFe for
Stations along the SEA, SEG and YBC lines are shown in Figures 3-6. In general, DFe, LPFe
and RPFe exhibited similar trends with depth (Figures 7a-f).

23 Data from the SEA line are shown in Figure 3a-d. Station SEA20 (1011 m depth) is ~25 km

offshore from SEA5 (145 m depth). Within the surface mixed layer, waters were slightly warmer

and saltier at SEA20 (7.1 °C, S ~ 32.1) as compared to SEA5 (6.7 °C, S ~31.9) and Fe

26 decreased from inshore to offshore from ~ 1 nM DFe, ~ 2.6 nM LPFe, and ~ 6 nM RPFe at

27 SEA5 to ~ 0.5nM DFe, ~0.6 nM LPFe, and ~ 2 nM RPFe at SEA20. At Station SEA5, the

concentration of all Fe fractions increased with depth (2.39 nM DFe, 5.87 nM LPFe, and 15.6

nM RPFe) to roughly twice their surface value. Increases with depth at SEA20 resulted in

30 maximum concentrations of 2.58 nM LPFe, and 9.81 nM RPFe at 750 m. The maximum DFe at

31 SEA20 was 1.70 nM at 500m, decreasing to 1.17 nM at 750 m. An increase in light attenuation

(decrease in % Beam c) was observed at around 150 – 350 m, suggesting advection of sediment
 from the shelf. At these stations DFe accounted for ~ 17% to 50% of the reactive Fe pool.

3 Data from stations along the SEG line are presented in Figures 4a-d (first occupation) and 5ad (second occupation). Water depth at Stations SEG0 and SEG20 were 110 m and 1700 m, 4 respectively. These stations are ~38 km apart. Due to time constrains SEG20 was only profiled 5 down to 360 m on the second occupation. The surface mixed layer at Station SEG0 had shoaled, 6 7 warmed and freshened on the second occupation (Figures 4a and 5a), and exhibited an order of magnitude increase in Chl. The higher biomass in this fresher surface layer was likely advected 8 from inshore, as the large increase in Chl values were not apparent at Station SEG20 (Figures 4b 9 and 5b). At Station SEG20 salinity increased within the upper 250 m on the second occupation 10 suggest advection of offshore water or upwelling. At Station SEG0 LPFe and RPFe were low (~1 11 12 to 4 nM RPFe and 0.4 to 1.8 nM LPFe) and increased gradually with depth during both sampling periods. The concentration of DFe also tended to increase with depth and ranged from 0.5 to 1.3 13 14 nM. Offshore, at SEG20 DFe had a similar concentration range (0.5 to 1.17 nM) with the higher values at and below 200 m. Similar to SEA20, an increase in light attenuation was also observed 15 16 at around 150 to 350 m, but the change was gradual during the second occupation. The concentrations of LPFe and RPFe were higher offshore than over the shelf (SEG0), and ranged 17 18 from 0.6 to 3.6 nM for LPFe and from ~ 2 to 11 nM for RPFe. The concentration of DFe accounted for $\sim 20\%$ to 70% of the reactive Fe at these stations. 19 20 Data from Stations YBC10, YBC40 and YBC50+ are presented on Figures 6a-f. The shelf widens in this region and YBC10 (165m depth) is ~ 55 km inshore of YBC40 (190m depth), 21 22 which is ~ 30 km inshore of YBC50+ (1400 m depth). The surface mixed layer was slightly cooler and saltier offshore (7.0 °C, S ~ 32.4) as compared to the inshore station (7.4 °C, S ~32.1). 23 A bottom nepheloid layer was apparent from the reduction in % Beam c at Station YBC10 and to 24 a lesser extent at YBC40. An increase in light attenuation was also apparent from 50 to 100 m at 25 YBC 40 without a concomitant increase in Chl values. Similar to offshore stations SEA20 and 26 27 SEG20, Station YBC50+ exhibited an increase in light attenuation from 150 to 350 m. The 28 concentration of DFe at these stations tended to increase with depth. At YBC10 Fe was relatively 29 homogeneous ranging from 0.65 nM to 0.83 nM in DFe, from 0.38 nM to 0.77 nM in LPFe, and 30 with RPFe at ~ 1 nM throughout the water column (nepheloid layer not sampled). Surface Fe concentrations at Station YBC40 were similar to YBC10, and increased to 2.12 nM DFe, 18.9 31

nM LPFe, and 48.8 nM RPFe at 75 m, where light attenuation increased. AT YBC 50+, DFe 1 exhibited a minimum concentration of 0.30 nM at 50 m, within a tongue of colder (5.5 °C) and 2 higher salinity (32.5) water compared to the surface mixed layer, and had a maximum of 1.65 3 4 nM at 300 and 500m, decreasing to 1.24 nM at 1000 m. Here the concentrations of LPFe and 5 RPFe were highest at the surface (4.89 nM and 79.0 nM, respectively), lowest also at 50 m (0.79 6 nM and 3.54 nM, respectively), and increased in concentration at 300 m (2.16 nM and 12. 2nM, respectively) within the area where light attenuation was enhanced. The concentration of DFe 7 8 accounted for $\sim 10\%$ to 60% of the reactive Fe at these stations.

Two classes of organic Fe-binding ligands were determined in most samples collected at 20 9 m, though only one ligand class was present at SEA20 and YBC50+. Ligands in the 20 m spring 10 samples were generally weaker, with lower conditional stability constants, than measured in 11 spring surface (1 m) waters (Table 3). In these 20 m samples, the ligands determined were 12 13 classified across three ligand classes based on conditional stability constants (Table 3), though even the stronger L₁-type ligands were on the low end of conditional stability constants for this 14 ligand class (log $K_{FeL_i,Fe}^{cond} = 12.04 - 12.15$) compared to those measured within the mixed layer 15 along surface transects (Table 3; 12.16 - 13.0). 16

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18 3.3 Summer Surface Transects

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20 Surface temperature along the Seward Line was relatively warm and homogeneous, ranging from 10.4 °C to 11.0 °C (Figure 8a). In Prince William Sound, T was slightly warmer (11.4 °C), 21 and S was the freshest measured during the cruise (S = 24.9) (data not shown). Over the 22 23 continental shelf, salinity ranged from 26.1 at GAK 1 to 32.6 at GAK 13 (Figure 8b). Temperature and salinity were not obtained on 16 September when the MultiNet was not 24 25 deployed. Daily variability in N+N, DFe, and TDFe (Figures 7a-c) was observed along the Seward Line on re-occupied stations (GAK 4-7). The concentration of N+N from GAK 1 to 26 GAK 12 ranged between ~ 2 and 4 μ M on 17 and 19 September, and was elevated (9.92 μ M) at 27 28 the furthest offshore station, GAK 13 (Figure 8a), where Alaska Gyre water was sampled. Earlier (16 Sep), higher concentrations of N+N (~ 6-7 µM) were observed from GAK 4 to GAK 5.5, 29

but similar concentrations of $\sim 4 \,\mu\text{M}$ were observed from GAK 6 to GAK 7.5 (Figure 8a). In 1 Prince William Sound N+N was depleted (0.33 μ M) (data not shown). 2 The concentration of DFe ranged by two orders of magnitude from 0.052 nM at GAK 13 to 3 4.87 nM at GAK 1 (Figure 8b), and correlated to salinity, with higher DFe at lower salinity 4 values. Subnanomolar [DFe] were observed from GAK 4 to GAK 7.5 on 16 September, and 5 6 from GAK 5 to GAK 13 on 17 and 19 September (Figure 8b). The concentration of TDFe varied 7 by three orders of magnitude from 0.19 nM at GAK 13 to 585 nM at GAK1 (Figure 8c). From GAK 1 to GAK 12, TDFe was dominated by the ALPFe fraction, which was indicated by the 8 low percentage and relatively uniform (~ 1- 8%) fraction of DFe observed (Figure 8c). This 9 fraction increased to ~ 28% at GAK 13, where TDFe was found at subnanomolar concentrations. 10 11 Two classes of Fe-binding organic ligands were also determined in summer surface samples, except at GAK 9, where one ligand class was observed. As seen in the spring samples, the two 12 13 ligand classes observed fit the definitions of three ligand classes based on conditional stability 14 constants, though the stronger L₁-type ligands observed in summer presented intermediate conditional stability constants (log $K_{FeL_i,Fe^i}^{cond} = 12.0 - 12.4$) to those measured between the surface 15 and 20 m during spring (Table 3) in southeast Alaska. 16

17

18 4. Discussion

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20 4.1 Seasonal and spatial variability in reactive iron

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22 Reactive Fe, the sum of DFe and the labile fraction of suspended particles, constitutes the 23 pool of Fe that is potentially available for biological uptake. Field and laboratory studies (e.g. Chase et al., 2005; Hurst and Bruland, 2007; Rich and Morel, 1990) have shown that a fraction 24 25 of particulate Fe can be readily solubilized and become available for biological uptake on the time scale of days. Given that in the Gulf of Alaska coastal rivers provide a massive input of 26 27 particulate material (0.3-0.45 km³/yr; Menard, 1979), and roughly one third of the mountainous coastline here is irregularly covered by glaciers (Wang et al., 2004), glacially derived sediment is 28 likely the major source of reactive Fe to the coastal GoA (Lippiatt et al., 2010a). However, this 29 30 input is highly seasonal. The seasonal transport of sediment by coastal rivers peaks in late summer/early autumn (Milliman and Syvitski, 1992) when fresh water discharge is at its 31

maximum from snow melt and summer rains (Weingartner et al., 2005). Additional glacial 1 2 sediment input can happen from mid to late autumn, along both the southeast and western GoA 3 regions, during episodic gap wind events that promote dust storms when anomalous northerly 4 winds provide sufficient energy to lift and transport sediment from exposed glacial river beds (Ladd and Chen, 2015; Ladd et al., 2015; Crusius et al., 2011). Glacially derived sediment is 5 6 produced by physical weathering, and in coastal waters of the GoA, glacially derived suspended 7 particles have been shown to contain on average a lower percentage of LPFe (11% of suspended particulate Fe) as compared to suspended sediment from coastal waters influenced by watersheds 8 where chemical weathering dominates (e.g. Columbia River plume with 26% LPFe) (Lippiatt et 9 al., 2010b). Although glacially-derived suspended particles contain less labile Fe, the massive 10 input of particulate material along the GoA coast can result in highly elevated LPFe and [TDFe] 11 12 that reach micromolar (µM) levels in low salinity river plumes (Lippiatt et al., 2010a; Schroth et al., 2014). As a result, the relative contribution of DFe to the reactive Fe pool is smallest during 13 late summer, not only in coastal areas influenced by large glacial rivers such as the Copper and 14 Alsek rivers, but also in regions with smaller glacial river drainage. For example, during the late 15 summer of 2011, we observed highly elevated concentrations of ALPFe (580 nM) along the 16 17 Seward line in lower salinity (26.1) inshore water, and in surface shelf/shelf break waters along this transect the contribution of DFe (< 5 nM) to the reactive pool had a median of only $\sim 3\%$ 18 (Figure 8) in the late summer of 2011. 19

20 In contrast to the dominant contribution of particulate Fe to the reactive Fe pool observed in late summer, the partitioning of Fe showed different characteristics during spring 2011. Spring 21 sampling took place in the absence of large riverine particulate input (low river flow conditions) 22 23 and at a time when biomass was anomalously low (Strom and Fredrickson, 2015; Stabeno et al., 2015). Elevated ALPFe (> 100 nM) was sampled at the surface only in the vicinity of Yakutat 24 25 Bay and Kayak Island (Figure 2e-f), where the shelf broadens and glacial input becomes more prevalent, but the median [ALPFe] was 4.5 nM. Station samples from 20 m depth contained 26 27 relatively low concentrations of LPFe (~19 nM or less; median 1.06 nM). Although the fraction 28 of the suspended particulate Fe that was leachable varied (6-42% LPFe), on average ($\sim 25\%$) it 29 was more akin to the average % LPFe found in suspended sediment of waters influenced by the Columbia River plume (Lippiatt et al., 2010b). The contribution of DFe to the reactive Fe pool 30 tended to be greater during spring 2011 as compared to the late summer, with an average of 24% 31

in surface transects (1 m) and 41% in station profiles (20 m). Figure 9 shows DFe as a function 1 2 of reactive Fe concentrations. Within the lower range of reactive Fe values ($\sim 1-50$ nM), the 3 concentration of DFe tended to be higher during the spring than during late summer in 2011. 4 Also, data collected during the late summer of 2007 (Lippiatt et al., 2010a) was shifted towards higher reactive Fe concentrations, but similar DFe as seen in spring 2011 (Figure 9). The 5 difference in the partitioning of reactive Fe observed between the two seasons is likely driven by 6 7 a combination of differences in river discharge and biological uptake histories leading to the time of sampling. Nutrients were available (Figures 2d) in surface waters during spring 2011, and the 8 low biomass observed was the result of a delayed and greatly reduced spring bloom (Strom and 9 Fredrickson, 2015; Stabeno et al., 2015; Waite and Mueter, 2013), suggesting that biological 10 uptake had not yet depleted the DFe pool accumulated from the previous winter. Alternatively, 11 the differences in Fe partitioning observed during the two sampling seasons in 2011 could reflect 12 geographical differences, as our spring sampling locations were mainly south of the outflow of 13 14 larger rivers such as the Alsek and Copper rivers (Figure 1). Data from late summer 2007 (Lippiatt et al. 2010a) also point to geographical differences in reactive Fe partitioning between 15 16 the southeast GoA and the Seward Line, supporting the notion that differences in reactive Fe observed in 2011 might reflect geographical, in addition to temporal differences (Figure 9). 17 18 Due to the conditions encountered, the partitioning of reactive Fe observed during May 2011 is likely more representative of that found during early spring in the southeast GoA, when river 19 20 runoff and the spring bloom have yet to get underway. In late winter, when prevailing winds stimulate downwelling (Stabeno et al., 2004; Weingartner et al., 2005), and fluvial and aeolian 21 22 (Crusius et al., 2011) inputs are essentially absent, the main mechanism by which reactive Fe is supplied to the surface is deep mixing by storms. The mixed DFe is expected to persist into 23 24 spring more readily than resuspended sedimentary particulate Fe. Thus, conditions leading into 25 the spring bloom in the Gulf of Alaska are likely characterized by a reactive Fe pool with a more dominant DFe contribution. Similarly, coastal water entrained into the ubiquitous mesoscale 26 eddies of the GoA at the time of their formation over the southeast GoA shelf, which takes place 27 during late winter/early spring (Crawford, 2002; Ladd et al., 2007; Henson and Thomas, 2008), 28 29 is expected to have a reactive Fe pool not heavily dominated by labile particulate Fe. 30

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1 4.2 The fate of the glacial Fe input

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3 Glacial input of iron into the Gulf of Alaska varies greatly with season, with the majority of the input taking place during summer via fluvial input, although during fall episodic offshore 4 flowing gap wind events (Ladd and Chen, 2015; Ladd et al., 2015) can bring glacial dust to the 5 6 Gulf of Alaska via atmospheric deposition. The bulk of the riverine input is trapped within the 7 buoyancy driven ACC, and freshwater plumes rich in dissolved and particulate Fe greatly reduce their Fe load as they mix with coastal waters, which tend to carry similar concentrations of DFe 8 9 in late summer (Lippiatt et al., 2010a, and this study) as in spring (this study). The availability of dissolved Fe-binding ligands likely caps the concentration of DFe as suggested by Lippiatt et al. 10 (2010a) and observed during this study. Sharp, non-conservative decreases in Fe concentrations 11 12 (of several orders of magnitude) across salinity gradients have been observed in late summer along southeast and northern coastal Alaska (Wu et al., 2009; Lippiatt et al., 2010a; Schroth et 13 14 al., 2014) that indicate the rapid loss of dissolved and suspended Fe loads from freshwater plumes along the coastally trapped ACC. Yet, even after rapid and substantial removal, the 15 16 enhanced input of glacial sediment in late summer results in labile particulate Fe disproportionately contributing to and augmenting the reactive Fe pool of the summer ACC. The 17 18 summer ACC represents a reservoir of reactive Fe and provides a vehicle for alongshore transport of Fe (Wu et al., 2009). The fate of the bulk of the glacially-derived iron is to be 19 20 deposited to the sediment, where it will be subsequently subjected to sedimentary transformation, and could ultimately contribute to the transport of Fe from the shelf to the interior via 21 22 mechanisms of subsurface lateral advection (Lam et al., 2006; Cullen et al., 2009). Data from our 23 vertical profiles support the idea that suspended particles from the shelf could be laterally 24 transported offshore contributing Fe to the subsurface. Bottom nepheloid layers were observed at 25 the wider shelf stations YBC10 and YBC40 (Fig. 6a-d), and at station YBC50+ the reduction in Beam c at ~150-250 m (Fig. 6f) suggest particles advection at these intermediate depths. 26 27 Although the differences in Beam c at these depths compared to 50 m immediately above and below are small, the difference is statistically significant (Mann-Whitney Rank Sum Test; P = <28 29 0.001). Particle advection is less apparent at stations SEA20 and SEG20 where the shelf is quite narrow. However, Beam c data at intermediate depths are significantly different than those from 30 31 layers immediately above and below.

2 4.3 Iron-binding Organic Ligands and DFe

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4 The correlation between DFe and the reactive Fe fraction tends to be variable and non-linear (e.g. Figure 9, and Aguilar-Islas et al., 2013), and although Fe can exchange between the labile 5 6 particulate and dissolved phases at seasonal and shorter timescales, the amount of DFe in 7 seawater has been shown to rarely exceed the concentration of the strong Fe(III)-binding organic ligand class (L₁-type) (Buck et al., 2007; Biller et al., 2013). For the fresh-water plumes of the 8 coastal GoA, Lippiatt et al. (2010a) speculated that low DOC in glacial melt water (Hood and 9 Scott, 2008) results in low concentrations of L_1 -type ligands, and that the low availability of 10 excess ligands caps [DFe] in summer plumes leading to the observed lower than expected [DFe] 11 12 relative to the highly elevated labile particulate Fe present in the system. We measured the concentration of dissolved Fe(III)-binding organic ligands during this study, and found that the 13 14 concentrations of the stronger L_1 -type ligands were in excess of DFe in all surface and subsurface spring, as well as in summer surface samples where an L_1 ligand class was identified 15 16 (Table 3; Figure 10). This finding supports the notion that the availability of strong Fe(III)binding organic ligands provides an upper bound for the concentration of DFe (Buck et al., 17 18 2007), and indicates that in GoA coastal waters L_1 -type ligands are not necessarily fully titrated by Fe. The concentration of L₁-type ligands does seem to cap the contribution of DFe to the 19 20 reactive Fe pool (Figure 10), and in absence of a greatly enhanced L₁-type ligand pool during late summer when particulate riverine input is high, the labile particulate fraction will tend to 21 22 become disproportionately dominant.

23 During both spring and summer 2011 DFe was strongly bound in surface waters. Conditional stability constants provide a measure of the binding strength of a given ligand class 24 with respect to inorganic Fe (Fe'), and we observed particularly high values of log $K_{FeL_i,Fet}^{cond}$ 25 26 (Table 3) in spring surface samples. These high conditional stability constants are consistent with 27 values determined for siderophores (Rue and Bruland, 1995; Buck et al., 2007; 2010), which are especially strong iron-binding ligands produced by bacteria for iron acquisition. They are also as 28 29 high or higher than the conditional stability constants determined for unknown ligands produced 30 by ambient biological communities in austral winter Southern Ocean coastal waters exposed to

spring light and temperature conditions (Buck et al., 2010). Thus, the higher log $K_{Fel_{ref}}^{cond}$ values 1 observed here, together with L₁ concentrations in excess of DFe are consistent with biological 2 ligand production in surface waters of the Gulf of Alaska during spring 2011. As previously 3 mentioned, spring 2011 had an anomalously reduced spring bloom (Stabeno et al., 2015), with 4 5 the phytoplankton community at the time of sampling consisting of low-light adapted cells 6 dominated by picophytoplankton ($< 2 \mu m$ cells) (Strom and Fredrickson, 2015). Although our 7 sampling took place in May, the characteristics of the Fe-binding ligand pool we observed in the spring of 2011 may be more akin to the ligand pool found in coastal GoA waters during early 8 9 spring prior to the onset of the spring bloom.

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11 4.4 Biological implications

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Diatom blooms contribute significantly to the productivity of the GoA in coastal and offshore 13 14 waters (Strom et al., 2006; Peterson and Harrison, 2012). The ability of diatoms to adjust their cellular Fe requirements and rapidly respond to changing available Fe concentrations plays a role 15 16 in their success. Diatoms can either lower their cellular Fe quotas as available Fe concentrations become extremely low (e.g. McKay et al., 1997), or they can perform "luxury uptake" under 17 conditions of high bioavailable Fe concentrations (Sunda and Huntsman 1995). Based on Fe 18 availability, intracellular Fe:C ratios vary widely in diatoms, with coastal species requiring 19 higher ratios than their oceanic counterparts (Sunda and Huntsman, 1995; Bruland et al., 2001; 20 Marchetti et al., 2006). Field and laboratory studies (e.g. Bruland et al., 2001; Sunda and 21 22 Huntsman, 1995; Marchetti et al., 2006) have shown Fe:C ratios in non-Fe limited coastal 23 diatoms can vary from about 20 µmol Fe: mol C at subnanomolar available Fe concentrations, to approximately 50 µmol:mol at ~1 nM available Fe, to over 100 µmol:mol at available Fe 24 concentrations > 3 nM. Although DFe is quantitatively chelated by organic Fe-binding ligands, 25 DFe provides a good measure of the Fe that is available to the phytoplankton community in time 26 scales of days (Bruland et al., 2001). Algal Fe requirements are also affected by light availability. 27 Low-light acclimation by phytoplankton requires an increase in Fe compounds involved in 28 photosynthesis (Falkowski et al., 1981), and because most intracellular Fe is used in these 29 compounds (Raven 1988), low-light acclimated phytoplankton, including diatoms, are 30

particularly susceptible to Fe availability. Even when diatom growth rates are depressed by lightlimitation, diffusion allows for maximum Fe uptake rates, but due to their larger cell size as
compared to oceanic species, coastal diatoms will experience diffusion limitation at [DFe] of a
few tenths of a nanomolar (Sunda and Huntsman, 1995). During our spring survey, surface DFe
(0.28 nM – 4.91 nM) had a median concentration of 1.25 nM. These concentrations are well
above concentrations likely to result in diffusion limitation for coastal diatoms, and in the rage of
values for possible Fe "luxury uptake" by diatoms.

It is of interest to consider the macro- and micro-nutrient relationship of surface waters just 8 prior to the onset of the spring bloom in the GoA. Winter mixing brings high concentrations of 9 macronutrients into surface waters. Yet, due to particle scavenging of DFe, [DFe] mixed into the 10 surface are potentially not sufficient relative to nitrate concentrations, especially in the absence 11 12 of external DFe sources such as fluvial input. The conditions we encounter during May 2011 (low biomass and low-light adapted community dominated by picophytoplankton (Strom and 13 14 Fredrickson, 2015), reduced fresh water input conditions that suggest minor fluvial Fe input (USGS river flow data), and relatively high surface [N+N]) are more akin to early spring 15 16 conditions, prior to the onset of the phytoplankton bloom, and we suggest early bloom nutrient dynamics can be investigated using this data set. 17

Here our spring nutrient data is used to investigate whether DFe was sufficient relative to available nitrate to support a coastal diatom bloom during early spring in the absence of additional DFe input from fluvial or other sources. During May 2011, the [N+N] varied over two orders of magnitude (0.18 -12.27 μ M) and had a median concentration of 8.04 μ M. The excess Fe (Fe_(ex)) relative to N+N is calculated in terms of cellular Fe:C ratios.

23

 $Fe_{(ex)} = [DFe] - ([N+N] \times C:N \times Fe:C / 1000)$

Where the ratio of Fe:C is dependent on *in-situ* [DFe], and the units for [DFe] are nM, while 24 the units of [N+N] are μM . Given ambient [DFe] at this time (median = 1.25 nM), an Fe:C ratio 25 26 of ~50 µmol:mol would be an appropriate estimate for the Fe requirement of coastal diatoms over the GoA shelf during spring 2011. We calculate Fe(ex) (Table 4) using this ratio, and a C:N 27 28 Redfield ratio of 106:16. Our small surface data set suggests that, in the absence of additional Fe inputs, a spring bloom of coastal diatoms would eventually run out of Fe (negative Fe(ex)) over 29 30 most of the southeast shelf prior to becoming nitrate limited. Yet, the coastal GoA becomes limited by nitrate as the growing season progresses (Childers et al., 2005), indicating other 31

nutrients (including Fe) are found in excess of nitrate. As shown by previous late summer 1 2 observations, surface [DFe] can remain relatively elevated (~ 2 nM) in surface waters of coastal 3 southeast Alaska when nitrate concentrations are depleted (Lippiat et al., 2010a). This underscores the importance in the timing of external DFe delivery from fluvial (or other) 4 sources, and its along-shelf transport to the productivity of the coastal GoA. Suspended particles 5 from glacial rivers carry a reservoir of reactive Fe that can be readily mobilized into the 6 7 dissolved phase as needed throughout the growing season. Interestingly, although the median surface [DFe] (1.25 nM) in May 2011 was not indicative of Fe limitation, other parameters 8 discussed above suggest a potential role for Fe in the productivity of spring 2011. Namely, the 9 pool of the of strong L₁-class Fe-binding ligand being consistent with *in-situ* production (Section 10 4.3), and the relatively high availability of N+N in surface waters along with the low diatom 11 abundance and low biomass (Stabeno et al., 2015; Strom and Fredrickson, 2015). 12 Summer DFe data (Wu et al., 2009; Lippiatt et al., 2010a; this study) along the Seward Line 13 show variable [DFe] as a function of surface salinity (Figure 11) that likely result from a 14 combination of inter-annual differences in biological uptake and the source of the surface water 15 16 traversing the region. In late summer 2011, surface offshore waters along the Seward Line exhibited low [DFe] that were deficient relative to available nitrate (even when considering a 17 lower Fe:C of 20 µmol:mol, appropriate given the subnanomolar DFe concentraions) (Table 4). 18 19 The Seward Line is located in an area highly impacted by the trajectory of Yakutat eddies (Ladd et al., 2007; Januot et al., 2009). The presence/absence of eddies and associated "streamers" 20 (Januot et al., 2009) likely contribute to the inter-annual variability of macronutrient and [DFe] 21 over the outer shelf and slope. Although variable, DFe appears to reach exceedingly low 22 23 concentrations of only a few tenths of a nanomolar or lower when surface salinity is > 32 (Figure 11). In this outer shelf/slope region, where we observed deficient [DFe], coastal diatoms may 24 25 experience diffusion limitation as [DFe] fall below a few tenths of a nanomolar. Thus, when higher salinity water (>32) from the oceanic GoA meanders over the shelf, the very low [DFe] 26 27 likely limit diatom production, and contribute to interannual biomass variability observed along the Seward line. 28

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30 5. Conclusions

Prior to the large riverine discharge that is received by the coastal GoA during summer, the 1 2 partitioning of reactive Fe over the southeast GoA shelf was in contrast to that described for late 3 summer (Lippiatt et al., 2010), and was characterized by suspended particles with higher %LPFe and a reactive Fe pool with a larger contribution by DFe as compared to previous summer 4 observations. Two classes of organic Fe-binding ligands were observed, with the stronger Fe-5 binding classes (L_1, L_2) found in excess of DFe. Excess L_1 concentrations with especially high 6 7 Fe-binding constants may suggest *in-situ* ligand production by an ambient biological community stressed by Fe availability at the surface in spring 2011. 8

9 Due to the reduced river discharge and biomass concentrations encountered during May 10 2011, our spring samples are likely more representative of late winter/early spring pre-bloom 11 conditions over the shelf. Nutrient characteristics of GoA waters prior to the growing season 12 have implications for interannual variation in productivity. Our samples provide a look at the 13 possible physico-chemical partitioning of Fe and its relation to macronutrients relevant for eddy 14 source waters and the initiation of the bloom. A better understanding of the magnitude of nutrient 15 fluxes and the mechanisms involved in nutrient dynamics over the GoA is needed.

16

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1 **References**

- Aguilar-Islas, A. M., Rember, R., Nishino, S., Kikuchi, T., and Itoh, M., 2013. Partitioning and
 lateral transport of iron in the Canada Basin. Polar Sci., doi:10.1016/j.polar.2012.11.001.
- Aguilar-Islas, A.M. and Bruland, K.W., 2006. Dissolved manganese and silicic acid in the
 Columbia River plume: A major source to the California Current and coastal waters off
 Washington and Oregon. Mar. Chem. 101, 233-247.
- Bell, J., Betts, J., and Boyle, E., 2002. MITESS: a moored in situ trace element serial sampler for
 deep-sea moorings. Deep-Sea Res. I 49, 2103-2118.
- 9 Berger, C.J.M., Lippiatt, S.M., Lawrence, M.G., Bruland, K.W., 2008. Application of a chemical
- 10 leach technique for estimating labile particulate aluminum, iron, and manganese in the Columbia
- River plume and coastal waters off Oregon and Washington. J. Geophys. Res.-Oceans 113,
 C00B01.
- Biller, D.V., Coale, T.H., Till, R.C., Smith, G.J., Bruland, K.W., 2013. Coastal iron and nitrate
- 14 distributions during the spring and summer upwelling season in the central California Current
- upwelling regime. Cont. Shelf Res. 66, 58-72
- 16 Boyd, P.W., Wong, C.S., Merrill, J., et al., 1998. Atmospheric iron supply and enhanced vertical
- carbon flux in the NE subarctic Pacific: Is there a connection? Global Biogeochem. Cycles 12,
 429-441.
- 19 Boyd, P.W., Law, C.S., Wong, C.S., Nojiri, Y., Tsuda, A., Levasseur, M., Takeda, S., Rivkin, R.,
- 20 Harrison, P.J., Strzepek, R., Gower, J., McKay, R.M., Abraham, E., Arychuk, M., Barwell-
- 21 Clarke, J., Crawford, W., Crawford, D., Hale, M., Harada, K., Johnson, K., Kiyosawa, H., Kudo,
- 22 I., Marchetti, A., Miller, W., Needoba, J., Nishioka, J., Ogawa, H., Page, J., Robert, M., Saito,
- 23 H., Sastri, A., Sherry, N., Soutar, T., Sutherland, N., Taira, Y., Whitney, F., Wong, S.E.,
- 24 Yoshimura, T., 2004. The decline and fate of an iron-induced subarctic phytoplankton bloom.
- 25 Nature 428, 549-553.
- Brown, M. T., Lippiatt, S. M., Lohan, M. C., et al., 2012. Trace metal distributions within a Sitka
- eddy in the northern Gulf of Alaska. Limnol. Oceanogr. 57, 503-518.
- 28
- 29 Bruland, K. W., Rue, E. L., Smith, G. J., 2001. The influence of iron and macronutrients in coastal
- 30 upwelling regimes off central California: implications for extensive blooms of large diatoms.
- 31 Limnol. Oceanogr. 46, 1661–1674.
- 32
- 33 Bruland, K.W., Rue, E.L., Smith, G.J., DiTullio, G.R., 2005. Iron, macronutrients and diatom
- blooms in the Peru upwelling regime: brown and blue waters of Peru. Mar. Chem. 93, 81–103.
 Buck, K. N., M. C. Lohan, C. J. M. Berger, and K. W. Bruland. 2007. Dissolved iron speciation
- in two distinct river plumes and an estuary: Implications for riverine iron supply. Limnol.
- 37 Oceanogr. 52, 843-855.
- Buck, K.N., Lohan, M.C., Berger, C.J.M., Bruland, K.W., 2007. Dissolved iron speciation in two
- distinct river plumes and an estuary: Implications for riverine iron supply. Limnol. Oceanogr., 52,
- 40 843-855.

- 1 Buck, K. N., J. Moffett, J., K. A. Barbeau, K.A., et al., 2012. The organic complexation of iron
- and copper: an intercomparison of competitive ligand exchange-adsorptive cathodic stripping
 voltammetry (CLE-ACSV) techniques Limnol. Oceanogr.-Methods 10, 496-515
- Buck, K.N., Selph, K.E., Barbeau, K.A., 2010. Iron-binding ligand production and copper
 speciation in an incubation experiment of Antarctic Peninsula shelf waters from the Bransfield
 Strait, Southern Ocean. Mar. Chem. 122, 148-159.
- 7
- Bundy, R., D.V. Biller, K.N. Buck, K.W. Bruland, and K.A. Barbeau, 2014. Distinct pools of
 dissolved iron-binding ligands in the surface and benthic boundary layer of the California Current. *Limnology and Oceanography*, 59, 769-787.
- 11
- 12 Chase, Z., Hales, B., Cowles, T., Schwartz, R., van Geen, A., 2005. Distribution and variability of
- iron input to Oregon coastal waters during the upwelling season. Journal of Geophysical Research Oceans 110, C10S12
- 14 Oceans 110, C10S12.
- 15 Childers, A.R., Whitledge, T.E., Stockwell, D.A., 2005. Seasonal and interannual variability in
- the distribution of nutrients and chlorophyll a across the Gulf of Alaska shelf: 1998-2000. Deep-
- 17 Sea Res. ll 52, 193-216.
- 18 Christensen, H.H., Mastrantonio, L., Gordon, J.C., Bormann, B.T., 2000. Alaska's Copper River:
- humankind in a changing world. United States Department of Agriculture, General Technical
 Benort DNW CTP 480, 20 p
- 20 Report, PNW-GTR-480, 20 p.
- 21 Crawford, W.R., 2002. Physical characteristics of Haida eddies. J. Oceanogr. 58, 703–713.
- 22 Crusius, J., Schroth, A. W., Santiago, G. et al., 2011. Glacial flour dust storms in the Gulf of
- 23 Alaska: Hydrologic and meteorological controls and their importance as a source of bioavailable
- 24 iron. Geophys. Res. Lett. 38, 10.1029/2010GL046573
- Cullen, J. T., Chong, M., Ianson, D., 2009. British Columbian continental shelf as a source of
 dissolved iron to the subarctic northeast Pacific Ocean. Global Biogeochem. Cycles 23, GB4012
- Elmwood, M.J., van der Berg, C.M.G., 2001. Determination of organic complexation of cobalt in
 seawater by cathodic stripping voltammetry. Mar. Chem 75, 33-47.
- 29
- 30 Falkowski, P.G., Owens, T.G., Ley, A.C., Mauzerall, D.C., 1981. Effects of growth irradiance
- levels on the ratio of reaction centers in two species of marine phytoplankton. Plant Physiology,
- 68, 969–973
- 33
- Gledhill, M., Buck, K.N., 2012. The organic complexation of iron in the marine environment: A
 review. Front. in Microbiol.: Microbiolog. Chem. 3, Article 69.
- 36 Gordon, L.I., J. C. Jennings Jr., A. A. Ross, J. M. Krest (1994). A suggested protocol for
- 37 continuous flow automated analysis of seawater nutrients (phosphate, nitrate, nitrite and silicic
- acid) in the WOCE Hydrographic Program and the Joint Global Ocean Fluxes Study. WHP
- 39 Operations and Methods. WOCE Hydrographic Program Office, Methods Manual 91-1.
- 40 November.

- 1 Henson, S.A., Thomas, A.C., 2008. A census of oceanic anticyclonic eddies in the Gulf of Alaska.
- 2 Deep-Sea Res. I 55, doi:10.1016/j.dsr.2007.11.005.
- 3 4
- Hood, E., Scott, D., 2008. Riverine organic matter and nutrients in southeast Alaska
- 5 affected by glacial coverage. Nature Geoscience 1, 583–587.
- 6
- Hurst, M.P., Bruland, K.W., 2007. An investigation into the exchange of iron and zinc between
 soluble, colloidal, and particulate size-fractions in shelf waters using low abundance isotopes as
- 9 tracers in shipboard incubation experiments. Marine Chemistry, 103, 211–226.
- 10
- 11 Hurst, M.P., Aguilar-Islas, A.M., Bruland, K.W., 2010. Iron in the southeastern Bering Sea:
- 12 elevated leachable particulate Fe in shelf bottom waters as an important source for
- 13 surface waters. Continental Shelf Res. 30, 467–480.
- 14
- Hutchins, D. A., G. R. DiTullio, Y. Zhang, K. W. Bruland, 1998. An iron limitation mosaic in the
 California upwelling regime, Limnol. Oceanogr., 43, 1037-1054.
- 17
- 18 Janout, M. A., Weingartner, T. J., Okkonen, S. R., Whitledge, T. E., and Musgrave, D. L., 2009.
- Some characteristics of Yakutat eddies propagating along the continental slope of the northern
 Gulf of Alaska, Deep-Sea Res. II 56, 2444–2459
- Johnson, K. S., F. P. Chavez, V. A. Elrod, S. E. Fitzwater, J. T. Pennington, K. R. Buck, and P. M.
- 22 Walz (2001), The annual cycle of iron and the biological response in central California coastal
- 23 waters, Geophys. Res. Lett. 28, 1247–1250, doi:10.1029/2000GL012433.
- Johnson, W. K., L. A. Miller, N. E. Sutherland, and C. S. Wong (2005), Iron transport by mesoscale
 Haida eddies in the Gulf of Alaska, Deep Sea Res. II 52, 933–953, doi:10.1016/j.dsr2.2004.08.017.
- Johnson, K. S., F. P. Chavez, and G. E. Friederich, Continental-shelf sediment as a primary source
 of iron for coastal phytoplankton, Nature 398, 697-700 1999.
- Ladd, C., Kachel, N. B., Mordy, C. W., Stabeno, P., 2005. Observations from a Yakutat Eddy in
 the northern Gulf of Alaska. J. Geophys. Res. 110, C03003.
- Ladd, C., Mordy, C.W., Kachel, N.B., Stabeno, P.J., 2007. Northern Gulf of Alaska eddies and
 associated anomalies. Deep Sea Res. I 54, 487–509.
- Ladd, C., Cheng, W., 2015. Characteristic of "gap winds" from Cross Sound, Alaska and their
 effects on regional oceanography. Deep-Sea Res II this issue.
- Ladd, C., Cheng, W., Salo, S., 2015. Gap Winds near Kodiak Island, Alaska and effects on regional
 oceanography. Deep-Sea Res II this issue.
- Lam, P. J., Bishop, J. K., Henning, B. C. C., Marcus, M. A., Waychunas, G. A., and Fung, I. Y.
- 37 2006. Wintertime phytoplankton bloom in the subarctic Pacific supported by continental margin iron Clobal Biogeochem Cycles 20 CB1006 doi:10.1020/2005CB002557
- iron. Global Biogeochem. Cycles, 20, GB1006, doi:10.1029/2005GB002557.

- Lam, P. J., and Bishop, J. K. B., 2008. The continental margin is a key source of iron to the HNLC
- 2 North Pacific Ocean. Geophys. Res. Lett. 35, 10.1029/2008GL033294
- Landry, M. R. et al., 1997 Iron and grazing constraints on primary production in the central
 equatorial Pacific: an Eq Pac synthesis. Limnol. Oceanogr. 4, 405-418.
- 5 Lippiatt, S. M., Lohan, M. C., and Bruland, K. W., 2010a. The distribution of reactive iron in
- northern Gulf of Alaska coastal waters. Mar. Chem. 121, 187-199.
- 8 Lippiatt, S.M., Brown, M.T., Lohan, M.C., Berger, C.J.M., Bruland, K.W., 2010b. Leachable
- particulate iron in the Columbia River, estuary, and near-field plume. Estuarine Coastal and Shelf
 Science 87, 33–42.
- 11
- Lippiatt, S. M., Brown, M. T., Lohan, M. C., et al., 2011. Reactive iron delivery to the Gulf of Alaska via a Kanai addy. Deep See Res. 158, 1001, 1102
- 13Alaska via a Kenai eddy. Deep-Sea Res. I 58, 1091-1102
- Liu, X.W., and Millero, F.J., 2002. The solubility of iron in seawater. Mar. Chem. 77, 43-54.
- 15 Maldonado, M. T., and N. M. Price, 2000. Nitrate regulation of Fe reduction and transport by Fe-
- 16 limited Thalassiosira oceanica. Limnol. Oceanogr. 45, 814-826.
- Maldonado, M. T., and N. M. Price. 1999. Utilization of iron bound to strong organic ligands by
 plankton communities in the subarctic Pacific Ocean. Deep-Sea Res. II 46: 2447-2473.
- 19 Marchetti, A., Maldonado, M. T., Lane, E. S. and Harrison, P. J., 2006. Iron Requirements of the
- 20 Pennate Diatom Pseudo-nitzschia: Comparison of Oceanic (High-Nitrate, Low-Chlorophyll
- 21 Waters) and Coastal Species. Limnol. Oceanogr. 51, 2092-2101.
- 22

22 Martin, J.,Gordon, R., Fitzwater, S., Broenkow, W., 1989. VERTEX: phytoplankton/iron studies in

- the Gulf of Alaska. Deep-Sea Res. 36,649–680.
- 25 McKay RML, Geider RJ, LaRoche J (1997) Physiological and biochemical response of the
- 26 photosynthetic apparatus of two marine diatoms to Fe stress. Plant Physiol. 114, 615–622.
- 27
- 28 Menard, H.W., 1979, Holocene Sediment Volume on the Northeast Gulf of Alaska Continental
- 29 Shelf and Sediment Contribution of Present Day Rivers: US Geol. Survey Marine Geology and
- Coastal hydrology Pap. 1150, p. 152.
- Milliman, J.D., Syvitski, J.P.M., 1992. Geomorphic/tectonic control of sediment discharge to the ocean: the importance of small mountainous rivers. Journal of Geology 100, 525–544.
- 52 Occan: the importance of shall mountainous rivers. Journal of Geology 100, 525-544.
- Moore, J.K., Doney, S.C., Kleypas, J.A., et al., 2002. An intermediate complexity marine
 ecosystem model for the global domain. Deep-Sea Res. II 49, 403-462.
- Morel, F. M. M., Rueter, J. G., Price, N.M., 1991. Iron nutrition of phytoplankton and its possible
- importance in the ecology of ocean regions with high nutrient and low biomass, Oceanography 4,
 56.61
- 37 56-61.

- 1 Morton, P.L., Landing, W.M., Hsu, S.C., Milne, A., Aguilar-Islas, A.M., Baker, A.R., Bowie, A.
- 2 R., Buck, C.S., Gao, Y., Gichuki, S., Hasings, M., Hatta, M., Johansen, A.M., Losno, R., Mead,
- 3 C., Patey, M.D., Swarr, Vandermark, G., A., Zamora, L.M., 2013. Methods for the sampling and
- 4 analysis of marine aerosols: Results from the 2008 GEOTRACES aerosol intercalibration study.
- 5 Limnology and Oceanography Methods. 11: 62-78, doi: 10.4319/lom.2013.11.62.
- 6 Musgrave, D.L., Weingartner, T.J., Royer, T.C., 1992. Circulation and hydrography in the
- 7 northwestern Gulf of Alaska. Deep-Sea Research 39, 1499–1519.
- Peterson, T. D., and Harrison, P. J., 2012. Diatom dynamics in a long-lived mesoscale eddy in the
 northeast subarctic Pacific Ocean. Deep-Sea Res. I 65, 157-170
- 10 Raven, J.A., 1988. The iron and molybdenum use efficiencies of plant growth with different
- energy, carbon and nitrogen sources. New Phytologist, 109, 279–287
- 12
- Rich, H.W., Morel, F.M.M., 1990. Availability of well-defined iron colloids to the marine diatom
 thalassiosira–weisflogii. Limnology and Oceanography, 35, 652–662.
- 15 Rijkenberg, M.J.A., Gerringa, L.J.A., Carolus, V.E., et al., 2006. Enhancement and inhibition of
- iron photoreduction by individual ligands in open ocean seawater. Geochim. Cosmochim. Acta70, 2790-2805.
- 18 Rue, E. L. and K.W. Bruland., 1995. Complexation of Iron (III) by natural organic ligands in the
- 19 Central North Pacific as determined by a new competitive ligand equilibration/adsorptive
- 20 cathodic stripping voltammetry method. Mar. Chem., 50, 117-138.
- Schroth A.W., Crusius, J., Hoyer, I., and Campbell R., 2014. Estuarine removal of glacial iron
 and implications for iron fluxes to the ocean. Geophys Res. Letters 41, 3951-3958.
- Stabeno, P.J., et al., 2004. Meteorology and oceanography of the Northern Gulf of Alaska.
 Continental Shelf Res. 24, 859–897.
- Stabeno, P.J., Bond, N.A., Kachel, N.B., Ladd, C., Mordy, C.W., Strom, S.L., 2015. Southeast
- Alaskan shelf from southern tip of Baranof Island to Kayak Island: Currents, mixing and chlorophyll₂, Deep-Sea Res II this issue
- 27 chlorophyll-a. Deep-Sea Res II this issue.
- 28 Strom, S.L., Olson, M.B., Macri, E.L., Mordy, C.W., 2006. Cross-shelf gradients in
- phytoplankton community structure, nutrient utilization, and growth rate in the coastal Gulf of
 Alaska. Mar. Ecol. Prog. Ser. 328, 75-92.
- 31 Strom S. L., and Fredrickson, K. A., 2015. Spring phytoplankton in the eastern coastal Gulf of
- 32 Alaska: Photosynthesis and production during high and low bloom years. Deep-Sea Res II, this issue.
- 33
- Sunda, W., Huntsman, S.A., 1995. Iron uptake and growth limitation in oceanic and coastal
- 35 phytoplankton. Marine Chemistry 50, 189–206.

- 1 Waite, J.N., Mueter, F.J., 2013. Spatial and temporal variability of chlorophyll-a concentrations
- 2 in the coastal Gulf of Alaska, 1998-2011, using cloud-free reconstructions of SeaWiFS and
- 3 MODIS-Aqua data. Prog Oceanogr 116, 179-192.
- 4 Wang, J., Jin, M., Musgrave, D. L., Ikeda, M., 2004. A hydrological digital elevation model for
- 5 freshwater discharge into the Gulf of Alaska, J. Geophys. Res., 109, C07009,
- 6 doi:10.1029/2002JC001430.
- 7 Weingartner, T. J., Danielson, S. L., Royer, T. C., 2005. Fresh water variability and
- 8 predictability in the Alaska Coastal Current. Deep-Sea Research II 52, 169–191.
- 9 Wu, J.F., and Boyle, E.A., 1998. Determination of iron in seawater by high-resolution isotope
- 10 dilution inductively coupled plasma mass spectrometry after Mg(OH)₂ coprecipitation. Anal.
- 11 Chim. Acta 367, 183-191.
- Wu, J., 2007. Determination of picomolar iron in seawater by double Mg(OH)₂ precipitation
 isotope dilution high resolution ICPMS, Mar. Chem., 103, 370–381.
- 14 Wu, J.F., Aguilar-Islas, A., Rember, R., Weingartner, T., Danielson, S., and Whitledge, T., 2009.
- Size-fractionated iron distribution on the northern Gulf of Alaska. Geophysical Research Letters36, L11606.
- 17
- 18
- 19
- 20
- 21

- 1 Table 1. Procedurally Defined Fe Fractions. Analyzed fractions are in grey boxes, and calculated
- 2 fractions are white boxes

| Fe Fraction | Abbreviation | Definition and Notes |
|----------------------------|--------------|--|
| Dissolved Fe | DFe | The fraction that passes through the filter membrane (< 0.4 μ m) or filter capsule (< 0.2 μ m). Includes colloids and is considered to be available to biology. |
| Particulate Fe | | The fraction collected on the filter membrane (> 0.4 μ m). It is the sum of LPFe and RPFe |
| Leachable particulate Fe | LPFe | The particulate fraction that is leached in 2 h with a solution of 25% acetic acid and 0.02 M NH ₂ OH • HCl after a short heating step. This fraction can be mobilized into the dissolved phase over short time scales. |
| Refractory particulate Fe | <u>RPFe</u> | The particulate fraction that is not leached with the above solution. This fraction is not considered accessible to biology over seasonal time scales. |
| Reactive Fe | | The fraction that is potentially available to biology in short timescales (days to weeks). (DFe + LPFe; Lippiat et al., (2010)) |
| Total dissolvable Fe | <u>TDFe</u> | The fraction in an unfiltered sample acidified to pH 1.8. for > 1 month. Includes DFe and is analogous to reactive Fe |
| Acid-labile particulate Fe | ALPFe | The fraction of particulate Fe that dissolves when an unfiltered sample is acidified to pH 1.8 (TDFe – DFe). It is analogous to LPFe. |
| Labile particulate Fe | | Either LPFe or ALPFe |

| Sampling Season | Sampling Location | Sampling Method | Measured (Derived) Fe Fraction |
|--------------------|--|-----------------|---------------------------------------|
| Spring | Southeast GoA Profiles at SEA, SEG, and YBC | UAF Vanes | DFe, LPFe, RPFe (PFe, Reactive Fe) |
| | Southeast GoA Surface Samples | Towed Fish | DFe, TDFe (ALPFe) |
| Late Summer | Western GoA surface samples along Seward Line, PWS | Towed Fish | DFe, TDFe (ALPFe) |

1 Table 2. Sampling Season, Location and Method

1

| Sample | $[L_1]$ | $\log K_1$ | $[L_2]$ | log K ₂ | $[L_3]$ | log K ₃ | $[eL_1]$ | [L _{total}] | [eLtot] |
|---------------|---------|------------|---------|--------------------|---------|--------------------|----------|-----------------------|---------|
| Spring ~ 1 m | (nM) | | (nM) | | (nM) | | (nM) | (nM) | (nM) |
| T01-01 (ChS) | 2.73 | 13.00 | 0.42 | 11.84 | | | 0.31 | 3.15 | 0.73 |
| T01-02 (ChS) | 2.60 | 12.38 | 0.66 | 11.45 | | | 0.40 | 3.26 | 1.06 |
| T02-01 (CS) | 3.12 | 12.67 | 0.76 | 11.47 | | | 0.23 | 3.88 | 0.99 |
| T02-03 (CS) | 3.65 | 12.68 | 0.29 | 11.85 | | | 0.32 | 3.94 | 0.62 |
| T02-05 (CS) | 2.59 | 12.27 | 1.21 | 11.25 | | | 0.65 | 3.80 | 1.85 |
| T03-03 (KIs) | 1.92 | 12.71 | 1.19 | 11.17 | | | 0.55 | 3.10 | 1.73 |
| T03-07 (KIs) | 1.41 | 12.17 | 0.34 | 11.26 | | | 0.49 | 1.74 | 0.83 |
| T04-08* (KIs) | 3.78 | 12.36 | 0.47 | 11.97 | | | 0.81 | 4.25 | 1.28 |
| T05-03 (YBC) | 1.17 | 12.16 | 1.05 | 11.37 | | | 0.67 | 2.22 | 1.72 |
| T07-09 (ChS) | 1.57 | 12.80 | 0.54 | 11.40 | | | 0.42 | 2.11 | 0.96 |
| Spring 20 m | | | | | | | | | |
| YBC-10 | 1.36 | 12.15 | 0.70 | 11.53 | | | 0.65 | 2.06 | 1.34 |
| YBC-40 | | | 1.38 | 11.92 | 1.36 | 10.82 | | 2.74 | 1.95 |
| YBC-50+ | | | 2.76 | 11.66 | | | | 2.76 | 1.94 |
| SEA 20 | | | 1.74 | 11.40 | | | | 1.74 | 1.19 |
| SEG 0 | 1.00 | 12.05 | 0.50 | 11.41 | | | 0.46 | 1.50 | 0.96 |
| SEG 0a | | | 0.79 | 11.57 | 0.85 | 10.65 | | 1.64 | 1.16 |
| SEG 20a | 1.18 | 12.04 | | | 4.14 | 10.38 | 0.67 | 5.32 | 4.81 |
| Late Summer | ~1 m | | | | | | | | |
| GAK 5 | 1.43 | 12.14 | 0.54 | 11.44 | | | 1.00 | 1.97 | 1.54 |
| GAK 9 | | | | | 1.39 | 10.93 | | 1.39 | 1.16 |
| GAK 10 | 0.70 | 12.39 | 0.41 | 11.31 | | | 0.39 | 1.11 | 0.81 |
| GAK 13 | | | 0.74 | 11.67 | 0.96 | 10.73 | | 1.70 | 1.65 |
| GAK 5 | 0.35 | 12.01 | | | 2.91 | 10.23 | 0.20 | 3.26 | 3.11 |
| GAK 1 | 5.97 | 12.30 | 1.06 | 11.63 | | | 1.11 | 7.03 | 2.17 |
| | | | | | | | | | |

Table 3. Organic Speciation of DFe. 2

3 Ligand concentrations and conditional stability constants are reported as the average of the results from two interpretation

4 techniques applied to the data from each sample. Ligand classes (L1, L2, L3) defined by conditional stability constants (log

 $K_{FeL_i,FeL}^{cond}$; abbreviated as log K_i for i = 1, 2, 3), where L₁: log K₁ = > 12, L₂: log K₂ = 11 - 12, L₃: log K₃ = 10 - 11. Excess L₁ 5

6 $([eL_1])$ and excess total ligand $([eL_{tot}])$ concentrations calculated from subtracting ambient dissolved iron concentrations from L_1

7 $([L_1])$ and total ligand $([L_{tot}] = [L_1] + [L_2] + [L_3])$ concentrations, respectively.

8 * Suspected contamination during sample collection

1 Table 4. Nutrient relationships

| Cruise | Sample | Longitude [deg_east] | Latitude [deg_north] | Depth [m] | DFe [nM] | N+N [umol/kg] | Fe(ex) [nM] | Fe:C [umol:mol] |
|--------|--------|-------------------------|-------------------------|--------------|----------|------------------|----------------|--------------------|
| 1TT11 | T01-01 | -134.79 | 56.19 | Surface | 2.41 | | | |
| 1TT11 | T01-02 | -134.89 | 56.26 | Surface | 2.21 | | | |
| 1TT11 | T01-03 | -134.97 | 56.32 | Surface | 1.36 | | | |
| 1TT11 | T02_01 | -137.68 | 58.23 | Surface | 2.89 | 11.65 | -0.97 | 50 |
| 1TT11 | T02_02 | -137.84 | 58.32 | Surface | 3.74 | 12.08 | -0.26 | 50 |
| 1TT11 | T02_03 | -137.91 | 58.37 | Surface | 3.32 | 10.80 | -0.25 | 50 |
| 1TT11 | T02_04 | -137.98 | 58.41 | Surface | 1.83 | 11.05 | -1.83 | 50 |
| 1TT11 | T02_05 | -138.05 | 58.46 | Surface | 1.94 | 11.24 | -1.78 | 50 |
| 1TT11 | T03_01 | -141.36 | 59.68 | Surface | 2.77 | | | |
| 1TT11 | T03_02 | -141.64 | 59.66 | Surface | 1.25 | | | |
| 1TT11 | T03_03 | -142.78 | 59.61 | Surface | 1.37 | | | |
| 1TT11 | T03_04 | -143.01 | 59.60 | Surface | (>10)* | 6.23 | | |
| 1TT11 | T03_05 | -143.20 | 59.59 | Surface | 1.41 | 5.41 | -0.38 | 50 |
| 1TT11 | T03_06 | -143.39 | 59.58 | Surface | 1.52 | 5.57 | -0.33 | 50 |
| 1TT11 | T03_07 | -143.58 | 59.57 | Surface | 0.91 | 6.85 | -1.36 | 50 |
| 1TT11 | T03_08 | -143.83 | 59.55 | Surface | 0.98 | 6.72 | -1.25 | 50 |
| 1TT11 | T04_01 | -145.00 | 59.74 | Surface | 4.50 | 4.73 | 2.93 | 50 |
| 1TT11 | T04_02 | -145.16 | 59.74 | Surface | (>10)* | 0.30 | | |
| 1TT11 | T04_03 | -145.32 | 59.74 | Surface | 1.23 | 0.21 | 1.16 | 50 |
| 1TT11 | T04_04 | -145.42 | 59.70 | Surface | 1.84 | 0.18 | 1.78 | 50 |
| 1TT11 | T04_05 | -145.43 | 59.63 | Surface | 1.67 | 0.21 | 1.60 | 50 |
| 1TT11 | T04_06 | -145.42 | 59.58 | Surface | 0.76 | 7.71 | -1.80 | 50 |
| 1TT11 | T04_07 | -145.43 | 59.51 | Surface | 0.59 | 8.06 | -2.08 | 50 |
| 1TT11 | T04_08 | -145.42 | 59.39 | Surface | (2.97)* | 7.98 | 0.33 | 50 |
| 1TT11 | T04_09 | -145.44 | 59.31 | Surface | 1.70 | 8.06 | -0.97 | 50 |
| 1TT11 | T05_01 | -145.12 | 59.25 | Surface | 1.14 | 0.27 | 1.05 | 50 |
| 1TT11 | T05_02 | -141.15 | 58.62 | Surface | 0.35 | 8.57 | -2.49 | 50 |
| 1TT11 | T05_03 | -140.82 | 58.58 | Surface | 0.49 | 8.12 | -2.20 | 50 |
| 1TT11 | T05_04 | -140.35 | 58.52 | Surface | 0.28 | 8.69 | -2.60 | 50 |
| 1TT11 | T05_05 | -139.84 | 58.44 | Surface | 0.44 | 8.58 | -2.40 | 50 |
| 1TT11 | T05_06 | -139.40 | 58.37 | Surface | 1.50 | 8.48 | -1.31 | 50 |
| 1TT11 | T05_07 | -138.92 | 58.27 | Surface | 1.94 | 9.80 | -1.31 | 50 |
| 1TT11 | T05_08 | -138.43 | 58.20 | Surface | 0.94 | 9.65 | -2.26 | 50 |
| 1TT11 | T05_09 | -137.92 | 58.14 | Surface | 4.91 | 10.28 | 1.51 | 50 |
| 1TT11 | T05_10 | -137.62 | 58.11 | Surface | 0.63 | 8.89 | -2.31 | 50 |
| 1TT11 | T06_01 | -136.96 | 57.85 | Surface | 1.12 | 4.66 | -0.42 | 50 |
| 1TT11 | T06_02 | -137.12 | 57.81 | Surface | 0.97 | 0.24 | 0.89 | 50 |
| 1TT11 | T06_04 | -137.03 | 57.62 | Surface | 1.51 | 0.22 | 1.44 | 50 |
| 1TT11 | T06_05 | -136.89 | 57.48 | Surface | 0.57 | 1.40 | 0.10 | 50 |

| 1TT11 | T06_06 | -136.78 | 57.38 | Surface | 0.35 | 7.76 | -2.22 | 50 |
|-------|--------|---------|-------|---------|------|-------|--------|----|
| 1TT11 | T06_07 | -136.66 | 57.26 | Surface | 0.78 | 8.44 | -2.01 | 50 |
| 1TT11 | T07_01 | -135.69 | 56.77 | Surface | 1.34 | 2.64 | 0.46 | 50 |
| 1TT11 | T07_02 | -135.64 | 56.69 | Surface | 2.33 | 4.41 | 0.87 | 50 |
| 1TT11 | T07_03 | -135.58 | 56.61 | Surface | 0.77 | 8.69 | -2.11 | 50 |
| 1TT11 | T07_04 | -135.52 | 56.52 | Surface | 1.32 | 8.01 | -1.33 | 50 |
| 1TT11 | T07_05 | -135.45 | 56.43 | Surface | 0.76 | 8.89 | -2.19 | 50 |
| 1TT11 | T07_06 | -135.38 | 56.33 | Surface | 0.94 | 11.31 | -2.80 | 50 |
| 1TT11 | T07_07 | -135.29 | 56.20 | Surface | 0.87 | 10.04 | -2.46 | 50 |
| 1TT11 | T07_08 | -135.21 | 56.11 | Surface | 0.99 | 11.44 | -2.80 | 50 |
| 1TT11 | T07_09 | -135.15 | 56.02 | Surface | 1.15 | 12.27 | -2.91 | 50 |
| 2Tx11 | PWS0 | -147.82 | 60.52 | Surface | 4.61 | 0.33 | 4.57 | 20 |
| 2Tx11 | GAK4 | -149.06 | 59.41 | Surface | 0.36 | 6.70 | -0.53 | 20 |
| 2Tx11 | GAK4.5 | -149.00 | 59.35 | Surface | 0.55 | 7.10 | -0.40 | 20 |
| 2Tx11 | GAK5 | -148.94 | 59.29 | Surface | 0.43 | 6.01 | -0.37 | 20 |
| 2Tx11 | GAK5.5 | -148.86 | 59.20 | Surface | 0.55 | 6.30 | -0.28 | 20 |
| 2Tx11 | GAK6 | -148.79 | 59.13 | Surface | 0.92 | 4.12 | 0.37 | 20 |
| 2Tx11 | GAK6.5 | -148.70 | 59.01 | Surface | 0.37 | 4.42 | -0.22 | 20 |
| 2Tx11 | GAK7 | -148.65 | 58.97 | Surface | 0.46 | 3.97 | -0.07 | 20 |
| 2Tx11 | GAK7.5 | -148.59 | 58.90 | Surface | 0.41 | 3.32 | -0.03 | 20 |
| 2Tx11 | GAK1 | -149.47 | 59.87 | Surface | 4.87 | 2.29 | 4.56 | 20 |
| 2Tx11 | GAK2 | -149.31 | 59.68 | Surface | 1.56 | 3.66 | 1.08 | 20 |
| 2Tx11 | GAK3 | -149.20 | 59.58 | Surface | 3.27 | 2.93 | 2.89 | 20 |
| 2Tx11 | GAK4 | -149.05 | 59.43 | Surface | 1.70 | 3.78 | 1.19 | 20 |
| 2Tx11 | GAK5 | -148.90 | 59.28 | Surface | 0.15 | 3.78 | -0.35 | 20 |
| 2Tx11 | GAK6 | -148.77 | 59.14 | Surface | 0.40 | 3.52 | -0.07 | 20 |
| 2Tx11 | GAK7 | -148.63 | 58.95 | Surface | 0.21 | 3.15 | -0.21 | 20 |
| 2Tx11 | GAK8 | -148.49 | 58.80 | Surface | 0.25 | 3.93 | -0.27 | 20 |
| 2Tx11 | GAK9 | -148.36 | 58.69 | Surface | 0.23 | 3.03 | -0.17 | 20 |
| 2Tx11 | GAK10 | -148.17 | 58.49 | Surface | 0.30 | 2.02 | 0.04 | 20 |
| 2Tx11 | GAK11 | -148.04 | 58.36 | Surface | 0.25 | 2.22 | -0.04 | 20 |
| 2Tx11 | GAK12 | -147.91 | 58.21 | Surface | 0.20 | 1.27 | 0.03 | 20 |
| 2Tx11 | GAK13 | -147.80 | 58.11 | Surface | 0.05 | 9.92 | -1.26 | 20 |
| 1TT11 | sea20 | 224.31 | 56.31 | 20.00 | 0.55 | 11.67 | -3.31 | 50 |
| 1TT11 | seg0 | 223.92 | 57.37 | 20.00 | 0.92 | 11.62 | -2.93 | 50 |
| 1TT11 | seg0 | 223.92 | 57.37 | 40.00 | 0.54 | 13.51 | -3.94 | 50 |
| 1TT11 | seg0 | 223.92 | 57.37 | 60.00 | 1.29 | 13.56 | -3.20 | 50 |
| 1TT11 | seg20 | 223.37 | 57.20 | 50.00 | 0.51 | 11.84 | -3.41 | 50 |
| 1TT11 | seg20 | 223.37 | 57.20 | 300.00 | 1.17 | 35.53 | -10.60 | 50 |
| 1TT11 | ybc50f | 218.93 | 58.78 | 20.00 | 0.82 | 9.11 | -2.20 | 50 |
| 1TT11 | ybc50f | 218.93 | 58.78 | 50.00 | 0.30 | 12.24 | -3.76 | 50 |
| 1TT11 | ybc40 | 219.38 | 58.94 | 20.00 | 0.79 | 9.12 | -2.23 | 50 |
| 1TT11 | ybc40 | 219.38 | 58.94 | 50.00 | 0.89 | 14.72 | -3.99 | 50 |

| 1TT11 | ybc40 | 219.38 | 58.94 | 75.00 | 2.12 | 18.89 | -4.14 | 50 |
|-------|--------|--------|-------|--------|------|-------|-------|----|
| 1TT11 | ybc10 | 219.94 | 59.34 | 20.00 | 0.71 | 10.01 | -2.60 | 50 |
| 1TT11 | ybc10 | 219.94 | 59.34 | 50.00 | 0.65 | 13.55 | -3.84 | 50 |
| 1TT11 | ybc10 | 219.94 | 59.34 | 75.00 | 0.83 | 14.90 | -4.11 | 50 |
| 1TT11 | seg20a | 223.40 | 57.20 | 20.00 | 0.51 | 13.38 | -3.92 | 50 |
| 1TT11 | seg20a | 223.40 | 57.20 | 50.00 | 0.99 | 19.53 | -5.48 | 50 |
| 1TT11 | seg20a | 223.40 | 57.20 | 75.00 | 0.61 | 19.90 | -5.98 | 50 |
| 1TT11 | seg20a | 223.40 | 57.20 | 100.00 | 0.51 | 13.66 | -4.01 | 50 |

| 2 | * 0 / 1 | | 1 . | 1 11 |
|---|-------------|----------------|--------|-------------------|
| / | * Suspected | contamination | during | sample collection |
| ~ | Duspected | containination | auring | sumple concetion |

1 Figure Captions

- 2 Fig. 1. Map showing the general surface circulation of the Gulf of Alaska, and locations sampled
- 3 for Fe and related parameters. White circles indicate locations where surface transects were
- 4 sampled. Black circles show stations where Fe profiles were collected. Heavy outlines on circles
- 5 denote locations where Fe-binding organic ligand samples were collected. The Alaska Coastal
- 6 Current is indicated by thin black arrows. Also labeled are Prince William Sound (PWS), Kayak
- 7 Island (K Is), Yakutat Bay (Y Bay), Cross Sound (CS), Chatham Strait (ChS), the mouths of the
- 8 Copper River (Copper R) and Alsek River (Alsek R). The 250 m contour line denotes the shelf
- 9 break.
- 10 Fig 2. Map of surface transects during spring 2011 along the southeast GoA shelf. Temperature
- 11 (a), salinity (b), and fluorescence (c) were obtained from the ship's underway system. Samples
- 12 for the analysis of nitrate plus nitrite (N+N) (d), DFe (e), and TDFe (f) were obtained from with
- 13 a trace-metal-clean towed sampler. Kayak Island (K Is), Yakutat Bay (Y Bay), Cross Sound (CS)
- 14 and Chatham Strait (ChS) are indicated along with the 250 m isobath.
- 15 Fig 3. Depth profiles at Station SEA5, located 5 km from shore, and SEA 20, located 20 km from
- shore. Temperature (a, c), salinity (a, c), % beam transmission (% Beam c) (b, d), and factory
- 17 calibrated Chl a (b, d) were collected from sensors on the ship's CTD rosette, and DFe (a, c),
- 18 LPFe (b,d) and RPFe (b, d) were collected from UAF vanes deployed above the CTD rosette.
- 19 Fig 4. Depth profiles from 7 May at Station SEG0, located < 1 km from shore, and SEG20,
- 20 located 20 km from shore. Temperature (a, c), salinity (a, c), % beam transmission (% Beam c)
- 21 (b, d), and factory calibrated Chl a (b, d) were collected from sensors on the ship's CTD rosette,
- and DFe (a, c), LPFe (b,d) and RPFe (b, d) were collected from UAF vanes deployed above the
- 23 CTD rosette.
- Fig 5. Depth profiles from 17 May at Station SEG0, located < 1 km from shore, and SEG20,
- located 20 km from shore. Temperature (a, c), salinity (a, c), % beam transmission (% Beam c)
- 26 (b, d), and factory calibrated Chl a (b, d) were collected from sensors on the ship's CTD rosette,
- and DFe (a, c), LPFe (b,d) and RPFe (b, d) were collected from UAF vanes deployed above the
- 28 CTD rosette.
- Fig 6. Depth profiles at Station YBC10, located 10 km from shore, YBC40, located 40 km from
- shore, and YBC50+, located > 50 km from shore. Temperature (a, c, e), salinity (a, c, e), % beam
- transmission (% Beam c) (b, d, f), and Chl a (b, d, f) were collected from the ship's CTD rosette,
- 32 and DFe (a, c, e), LPFe (b, d, f) and RPFe (b, d, f) were collected from UAF vanes deployed
- above the CTD rosette.
- Fig 7. Consolidated depth profiles at Stations SEA, SEG and YBC. DFe (a, d), LPFe (b, e), and
- RPFe (c, f) collected from UAF vanes deployed starting 10 m above the CTD rosette.

- 1 Fig. 8. Surface data during 16, 17, and 19 September 2011 along the Seward Line. a)
- 2 Temperature and N+N, b) Salinity and DFe, and c) TDFe and the % of reactive Fe in the
- 3 dissolved phase (% DFe).
- 4 Fig. 9. Scatterplot of DFe as a function of reactive Fe (TDFe or LPFe+DFe). Vertical dash line
- 5 marks 50 nM reactive Fe. Samples from this study (2011) were collected from the surface tow
- 6 fish during spring (white circles) and late summer (black triangles); samples from 2007 obtained
- 7 from Lippiatt et al. (2010a) and were also obtained with a surface tow fish.
- 8 Fig. 10. Scatterplot of DFe as a function of strong-Fe binding ligand concentrations $[L_1]$. The
- 9 dash line denotes the 1:1 line. Data is from 2011 samples collected from the surface tow fish in
- spring (grey circles) and summer (black circles), and from UAF vanes position at 20 m (white
- 11 triangles).
- 12 Fig. 11. Variability in DFe as a function of salinity along the Seward Line. Samples were
- 13 collected from GAK 1 to GAK 13 during this study (2011; black circles), 2007 (grey squares;
- Lippiatt et al., 2010a), and 2004 (white dimonds; Wu et al., 2009)









2 Figure 2





2 Figure 3 a-b





2 Figure 3 c-d





2 Figure 4 a-b















2 Figure 5 c-d





2 Figure 6 a-b











2 Figure 6 e-f









2 Figure 8

















